



Estimation of soil losses due to water erosion in the Dagua River Basin, Colombia

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ABSTRACT

The Dagua river basin, in Colombia, is the most important source of water for the Valle del Cauca ecosystem, however, due to poor agricultural practices, it has been affected by water erosion. This study aimed at estimating soil erosion in the Dagua river basin, using the universal soil loss equation (USLE). The results show that most of the area presents erosivities that are between 1000-5000 MJ.mm.ha⁻¹, corresponding to low and very low categories. On the other hand, erodibility ranged from 0.143 to 0.842 t.ha.h. MJ⁻¹ mm⁻¹ ha⁻¹, which is framed in the categories from weak to extremely erodible, where the low to medium category predominates. Regarding soil losses due to erosion, it was found that more than 20% of each of the municipalities of Dagua, Restrepo, La Cumbre, and Vijes, showed high and very high erosion, particularly in the areas with bare soils and crops such as pineapple, contribute strongly, sometimes reaching over 1000 t ha⁻¹ yr⁻¹. Therefore, it is important to promote practices such as contour or contour planting, integrated crop cover management, land uses that integrate trees, and in more critical cases to consider ecological restoration processes.

Keywords: *Basin; Erodibility; Soil conservation; Soil degradation; USLE;*

Estimación de pérdidas de suelo causadas por la erosión del agua en la cuenca del Río Dagua, Colombia

RESUMEN

La cuenca del río Dagua, Colombia, es la fuente de agua más importante para el ecosistema del Valle del Cauca, sin embargo, debido a las malas prácticas agrícolas, se ha visto afectada por la erosión hídrica. El objetivo de este estudio fue estimar la erosión del suelo en la cuenca del río Dagua, utilizando la ecuación universal de pérdida de suelo (USLE). Los resultados muestran que la mayor parte del área presenta erosividades que se encuentran entre 1000-5000 MJ.mm.ha⁻¹, correspondientes a categorías bajas y muy bajas. Por otro lado, la erodabilidad osciló entre 0,143 y 0,842 t.ha.h. MJ⁻¹ mm⁻¹ ha⁻¹, lo que se enmarca en las categorías de débil a extremadamente erodable, donde predominó la categoría baja a media. En cuanto a las pérdidas de suelo por erosión, se encontró que más del 20% de cada uno de los municipios de Dagua, Restrepo, La Cumbre y Vijes, presentaron erosión alta y muy alta, y en particular las áreas con suelos desnudos y cultivos agrícolas como la piña, contribuyeron fuertemente, alcanzando en ocasiones más de 1000 t ha⁻¹ año⁻¹. Por lo tanto, es importante promover prácticas como la siembra en contorno o contorno, el manejo integrado de la cobertura de cultivos, los usos de la tierra que integren árboles y, en casos más críticos, considerar procesos de restauración ecológica.

Palabras clave: *Cuenca; Conservación de suelos; Degradación de suelos; Erodabilidad; USLE.*

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1. Introduction

Soil degradation by erosion is one of the most important problems affecting both natural ecosystems and agroecosystems worldwide (Panagos *et al.*, 2018; Borrelli *et al.*, 2020; Sharda *et al.*, 2021). This process consists of the gradual loss of the surface layer of the soil due to the action of wind or water (Borrelli *et al.*, 2017; Ma *et al.*, 2021), which has led to the reduction of its productive potential, generating negative effects not only environmental but also socioeconomic (Pandey *et al.*, 2021).

Erosion is a natural process that is caused by the combination of several factors such as steep slopes, climate, soil characteristics, type, and state of general cover. In turn, it is a process that is accelerated by the influence of anthropic activities (clean crops, planting in favor of the slope, extensive livestock, absence of conservation practices) (Amundson *et al.*, 2015; Abdulkareem *et al.*, 2019), and that is of great importance if one takes into account that the soil is a non-renewable natural resource.

Thus, world food production is threatened by soil erosion, a figure that reaches average values of $30 \text{ t ha}^{-1} \text{ year}^{-1}$, which leaves around 10 million hectares of unproductive farmland each year (Gachene *et al.*, 2020). It has been pointed out that this phenomenon can cause annual socioeconomic losses of approximately 40 billion dollars worldwide (Crosson, 1995; Oldeman *et al.*, 2017), and in environmental terms, likely, more than 90% of the soil of the planet earth deteriorates in 2050 (Bouma and Montanarella, 2016; Keesstra *et al.*, 2016). On the other hand, approximately 48% of soils have already been degraded in Colombia (IDEAM and UDCA, 2015).

According to the above, over the last decades, research has been carried out through experimentation and modeling of erosion, which allows for counteracting its effects (Ma *et al.*, 2021). In most experimental methods, soil erosion is calculated either in runoff plots under natural rainfall or under simulated conditions (El Kateb *et al.*, 2013; Anache *et al.*, 2017). However, erosion predictions can be carried out by taking advantage of mathematical models that express the efficiency of sediment transport from a hillside and a network of channels (Kirby and Morgan, 1980; Némétová *et al.*, 2020).

The prediction of erosion has a long history of more than seven decades (Bennett, 1939; Alewell *et al.*, 2019), where one of the most used models is the USLE, taking into account its high degree of flexibility, data accessibility, and comparability of results that allow adapting the model to almost all types of conditions in the world (Alewell *et al.*, 2019).

This equation was developed in the 1950s, by the United States Department of Agriculture and the Soil Conservation Service, and is one of the most important tools in conservation planning, helping to predict soil loss (Meinen and Robinson, 2021), even at the basin level when integrated with geographic information systems (GIS) (Mitasova *et al.*, 1996; Iroumé *et al.*, 2011).

In Colombia, the hydrographic basin of the Dagua River is the most important source of water for the Valle del Cauca ecosystem (Aguirre *et al.*, 2017). This region has more than eight life zones, which represents an enormous environmental value (Loaiza *et al.*, 2014). In this territory, the dominant covers are agricultural and silvopastoral, and only 2,867 ha (0.36%) are covered by native forests (IAVH, 2015) (Reina-Rodríguez *et al.*, 2016).

Despite the above, due to poor agricultural practices, ecosystems have been deteriorated (Daza *et al.*, 2012; Cardona *et al.*, 2014), since more than half of the territory presents processes of water erosion (Loaiza *et al.*, 2012). In this way, mass movements and the annual dumping of 250,000,000 kg t of sediments into Buenaventura Bay have been reported, causing enormous socioeconomic and environmental losses (Reyes *et al.*, 2010; Loaiza *et al.*, 2014).

The objective of the research was to estimate the loss of soil due to water erosion in the Dagua river basin, Colombia, using the empirical model of soil erosion (USLE), to obtain information for decision-making regarding the conservation strategies and resource management for the study area.

2. Materials and methods

2.1 Location of the study area

The hydrographic basin of the Dagua River is located on the western flank of the western mountain range, on the Valle del Cauca Pacific (González *et al.*, 2016) and is made up of the municipalities of Dagua, Buenaventura, La Cumbre, Restrepo, Vijes, Yotoco, and Calima Darien. Its coordinates include $3^{\circ} 53'13.2''$ to $3^{\circ} 27'39.6''$ of North Latitude and $76^{\circ} 57'21.6''$ to $76^{\circ} 40'29.99''$ of West Longitude. It has a total area of 1,422 km², of which about 98.2% correspond to the hillside area and 1.8% to the marine plain (CVC, 2007).

The average annual rainfall in the upper part of the basin is 900-1600 mm, while in the lower part it can reach 8000 mm (Aguirre *et al.*, 2017).

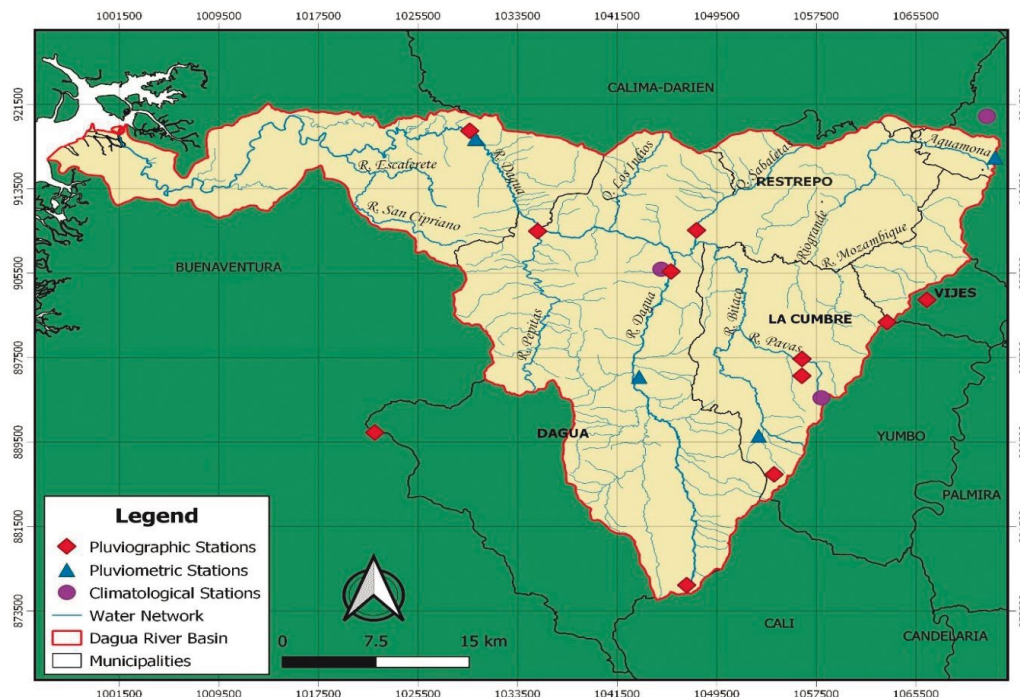


Figure 1. Location of the study area and the distribution of pluviometric, pluviographic and climatological stations.

2.2 Climatic data

To obtain climate information, records of eighteen climatological, pluviometric, and pluviographic stations belonging to the Autonomous Regional Corporation of Valle del Cauca (CVC), available for the study area, were analyzed, of which five are conventional (daily records) and thirteen are automatic (logs every 10 minutes); the above for a period of five years (2015-2019). Thus, Figure 1 presents this area, as well as the location of the stations used in the research.

2.3 Soil information

Based on the review of the information provided by the CVC advanced geographic viewer for the year 2021 and the document called “semi-detailed soil survey scale 1: 25000 of the basins prioritized by the CVC” from 2014, it was possible to obtain the cartographic information and identify the main soil orders present in the study area, which belong to the Inceptisols, Entisols, Histosols, Alfisols, Mollisols, and Andisols. Likewise, the properties required for the study were: texture, organic matter, permeability, and structure.

2.4 The Digital Elevation Model (DEM)

For the elaboration of the slope map, a digital elevation model supplied by the CVC with a resolution of 30 m was taken as input, with radiometric correction, from which it was possible to obtain the slope map. It was worked on a file geodatabase using Magna Sirgas coordinate system for Colombia.

2.5 Cover layer and land use

These layers were provided by the CVC (2021), which presents updated information for 2018. A total of 94 types of coverage were identified. The main covers correspond to the dense mixed highland forest, cultivated pasture, and secondary or transition vegetation, while the main established agricultural crop is pineapple (*Ananas comosus* (L.) Merr).

2.6 Erosion models

2.6.1 Universal Soil Loss Equation-USLE: The empirical USLE model was used, which has been disseminated and used worldwide for more than four decades (Alewell et al., 2019).

This model allows for estimating the long-term average erosion, and it is found in function of six factors, as presented in Equation 1 (Wischmeier and Smith, 1978).

$$A = \sum_{n=1}^6 RKLSCP \tag{1}$$

Where:

A is the soil loss calculated per unit area (t. ha. yr⁻¹), which is the result of the product of the six factors, R is the erosivity of rain, a factor that reflects its aggressiveness and runoff (MJ mm ha⁻¹ h⁻¹ yr⁻¹), K is the erodability, which represents the susceptibility to detachment of soil particles (t. ha. h. ha⁻¹. MJ⁻¹. mm⁻¹), L is a slope length factor (dimensionless), S is slope steepness (dimensionless), C is a plant cover factor (unitless), and P is a management practice factor.

R factor

To determine the R factor, precipitation events with a record greater than or equal to 12.7 mm, and separated by at least six hours (erosive events) were analyzed. This factor was calculated with the methodology proposed in the USLE by Wischmeier and Smith (1978); However, taking into account that in this case, the records of the pluviographic band are not available, but rather the accumulated precipitation in 10 minutes, it was proposed to use this interval to determine the average precipitation intensity (IM) required in the calculation of the kinetic energy (e) for each mm of rain, based on Equation 2:

$$e = 0.119 + 0.0873 * \log IM \tag{2}$$

Once the total energy was obtained in each 10-minute interval, the value of e was multiplied by the total precipitation recorded in said interval (Eq. 3):

$$Ei = e * PT \tag{3}$$

To obtain the total kinetic energy of the selected event (E), the energies obtained for each 10-minute interval were added and finally the equation used is the following:

$$R = \sum_{i=1}^m (EI_{30})_i \tag{4}$$

Where EI₃₀ is the erosivity of a single event (MJ mm ha⁻¹ h⁻¹), calculated as its total kinetic energy (E; MJ ha⁻¹) multiplied by the maximum rainfall intensity in 30 min (I₃₀; mm h⁻¹). The R factor was determined from the sum of the erosivity of each of the events considered in the year and an average value of the years considered for each station was then generated.

The R values obtained were classified according to the ranges proposed by Rivera and Gómez (1991).

Table 1. Classification of erosivity proposed by Rivera and Gómez (1991).

Class	Erosivity R (Mj. mm.h.ha ⁻¹ .yr ⁻¹)	Classification
1	<1000	Natural
2	1000-2500	Very low
3	2500-5000	Low
4	5000-7500	Moderate
5	7500-10000	High
6	10000-15000	Very high
7	15000-20000	Severe
8	> 20000	Extremely severe

K factor

The K factor was determined through the equation proposed by (Wischmeier and Smith, 1978) for soils containing less than 70% silt and very fine sand:

$$100 K = (2.1M^{1.14}(10^{-4})(12 - \alpha) + 3.25(b - 2) + 2.5 (c - 3) * 0.1317 \tag{5}$$

Where K is the erodibility of the soil (t ha h ha⁻¹ MJ⁻¹ mm⁻¹), M refers to the particle size (based on the soil texture), a is the percentage of organic matter, b is the soil structure code, and c is the kind of permeability. For those cases in which the organic matter content was greater than 4%, this value was assumed.

The K values obtained were classified according to the ranges proposed by USDA (1962).

Table 2. Classification of erodability proposed by USDA (1962).

K value (t. ha. h. MJ ⁻¹ mm ⁻¹ ha ⁻¹)	Classification
<0.0775	Very little erodable
0.0775- 0.1680	Weakly erodable
0.1680-0.3230	Moderately erodable
0.3230-0.6784	Strongly erodable
0.6784-1	Extremely erodable

LS factor

Given the rugged topography of the study area, in the case of factors L and S, the formula proposed by Pérez et al. (1995) was used:

$$LS = \left(\frac{\lambda}{22.6} \right)^{0.6} * \left(\frac{S}{9} \right)^{1.4} \tag{6}$$

Where LS is the slope length factor (unitless), λ is the slope length (m), and s is the slope expressed as a percentage.

Factors C and P

Factor C was determined by taking into account the information on the coverage consulted in the CVC advanced geographic viewer (2021). In the case of the P factor, considering that currently there is no report on the management of the coverage or the conservation practices that are carried out, it was assumed with a value of 1 (dimensionless), with which the highest value of the loss is calculated.

2.6.2 Factor zoning using GIS

The zoning of the USLE factors was carried out using the QGis version 3.6 software. In the case of factors K and C, the shapes obtained were converted to raster format. For the R factor, an interpolation was carried out through the deterministic method of the inverse of the distance (IDW), taking into account that it is a simple algorithm that is designed to reduce the distortion of the interpolation of the determined surface (Riquelme *et al.*, 2008), which is widely used in climate prediction and cartographic development (Núñez, 2014). Finally, the raster calculator was used to determine the LS factor and estimate soil losses due to erosion.

Results and discussion

3.1 Erosivity (R Factor)

The municipalities of Dagua, Restrepo, La Cumbre, and Vijes presented, according to the area, a rainfall erosivity mainly of a very low category (1000-2500 MJ.mm.ha⁻¹yr⁻¹) and low (2500-5000 MJ.mm.ha⁻¹yr⁻¹), while in Buenaventura there are mainly moderate and high categories, and sometimes even very high (5000-15000 MJ.mm.ha⁻¹yr⁻¹).

Of the eighteen stations evaluated, the two located in the municipality of Buenaventura presented high and very high values of pluvial erosivity (8753.76 MJ.mm.ha⁻¹yr⁻¹-10743.74 MJ.mm.ha⁻¹yr⁻¹), and the rest were among the categories very low, low and natural (636.68 MJ.mm.ha⁻¹yr⁻¹- 4383.98 MJ.mm.ha⁻¹yr⁻¹).

Figure 2 presents the behavior of rain erosivity, which shows an increase in R values as it approaches the western flank (higher areas of the basin). This behavior can be attributed to the fact that the R factor decreases as the altitude changes from the mountainous areas to the lowlands. This coincides with that indicated by Guauque *et al* (2021), who found the lowest values of the R factor in the highest areas of the Combeima basin (Colombian Andes), due to the low amount of rainfall.

Likewise, Riquetti *et al.*, (2020) modeled the R factor for South America, finding the lowest values of the R factor along the Andes Mountain Range due to the reduction in the amount of rainfall at the highest elevation, which shows that this factor can be influenced by the elevation of the mountainous regions on the continent.

There are no previous data reported for Valle del Cauca using this methodology. However, some investigations such as those carried out by (Ramirez *et al.*, 2009) (Castro *et al.*, 2017) in the departments of Caldas and Quindío, where the annual rainfall is between 1700 to 4000 mm, reporting erosivities classified as high and very high. On the other hand, a study carried out by Pacheco *et al.*, (2019) for an Ecuadorian tropical region, indicated that the ranges of erosivity are between 3162.12 and 12683.72 MJ.mm.ha⁻¹yr⁻¹.

3.2 Erodability (K factor)

There are six main soil orders within the Dagua river basin, with K factors ranging from 0.143 to 0.842 t.ha.h. MJ⁻¹mm⁻¹ha⁻¹, which falls into the weakly to extremely erodible categories. The order Inceptisol occupies the largest area with about 65.65% of the total land of the study area.

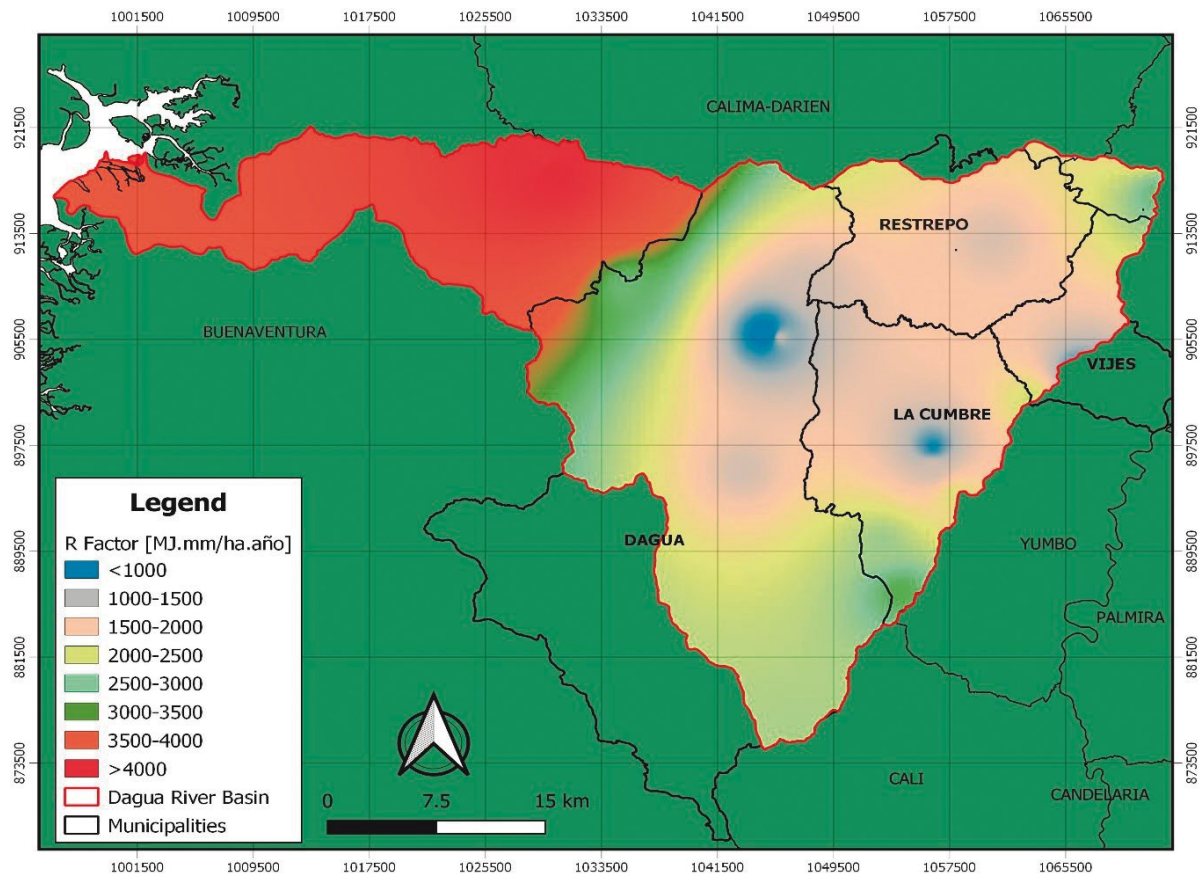


Figure 2. Rainfall erosivity in the Dagua river basin

At the municipal level, the highest values are found in Buenaventura and correspond to the Inceptisol order, which reaches 0.842 t.ha.h.MJ⁻¹mm⁻¹ha⁻¹. In second place, there is Dagua, with Inceptisol soils that reach values of up to 0.656 t.ha.h.MJ⁻¹mm⁻¹ha⁻¹. In the case of Vides and Yotoco, similar values are presented with maximums of 0.611 t.ha.h.MJ⁻¹mm⁻¹ha⁻¹ belonging to the order of molisols, while, in Restrepo and La Cumbre, maximums of up to 0.559 were reported, in some Andisols.

Regarding the minimum value of erodibility (0.143 t.ha.h.MJ⁻¹mm⁻¹ha⁻¹), this was the same for all municipalities. Figure 3 illustrates the K factor map in the study area.

Of the entire basin area, 86.88% correspond to the categories of weak to moderately erodible, 10.84% are strongly erodible and only 1.40% are extremely erodible.

The soil with the highest K factor value is dominated by very fine sand with silt particles, permeability, and low organic matter content, and they do not present a good structure, making them more susceptible to erosion. While the lowest K values presented high content of clay and organic matter, better structure, and higher permeability.

Few investigations have been carried out on erodability in the Dagua river basin. In this regard, Alarcón and Reyes (2013), in studies carried out in the La Centella micro-basin, Dagua, values of 0.003254 t.ha.h.MJ⁻¹mm⁻¹ha⁻¹ were found which are lower than those obtained in this research, taking into account that prior to the tests they supplied vermicompost, which modified the characteristics of the soil, due to the higher contents of organic matter, forming macroaggregates and improving the structure, increasing infiltration and reducing runoff, which influences the decrease in erodibility, despite this, it was highlighted that the erosion in the study area is high, especially considering the steep slopes and heavy rains.

3.3 Topographic factor (LS)

The highest values of this factor (five to seven, and higher than seven), extend from the northwest to the southwest, mainly covering the municipality of Dagua, and are attributed to steep slopes (12-75%) and even higher than 75%, corresponding to steep slopes to steep slopes with great lengths, which corresponds to the highest slope classification categories for Colombia according to IGAC (2014). This factor leads to a very powerful rain flow rate and aggravates soil erosion. Figure 4 presents the map of the LS factor in the study area.

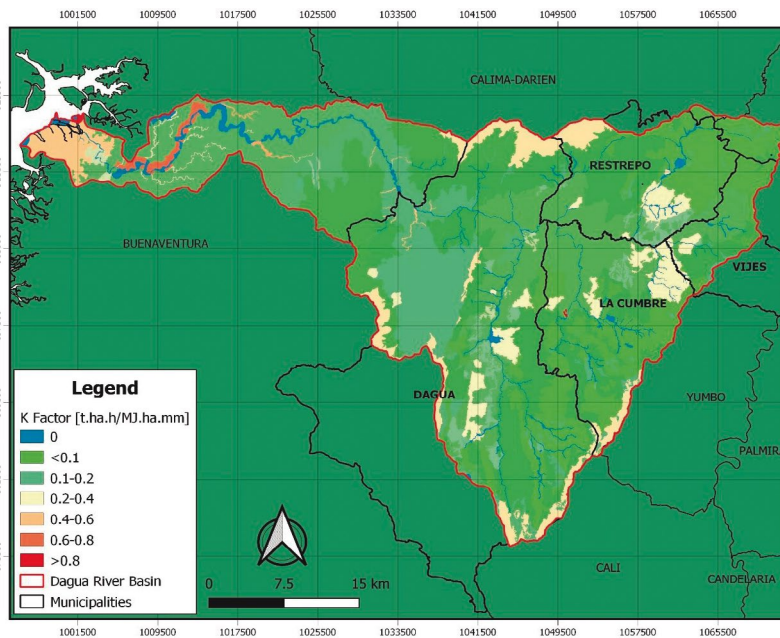


Figure 3. K factor in the Dagua river basin

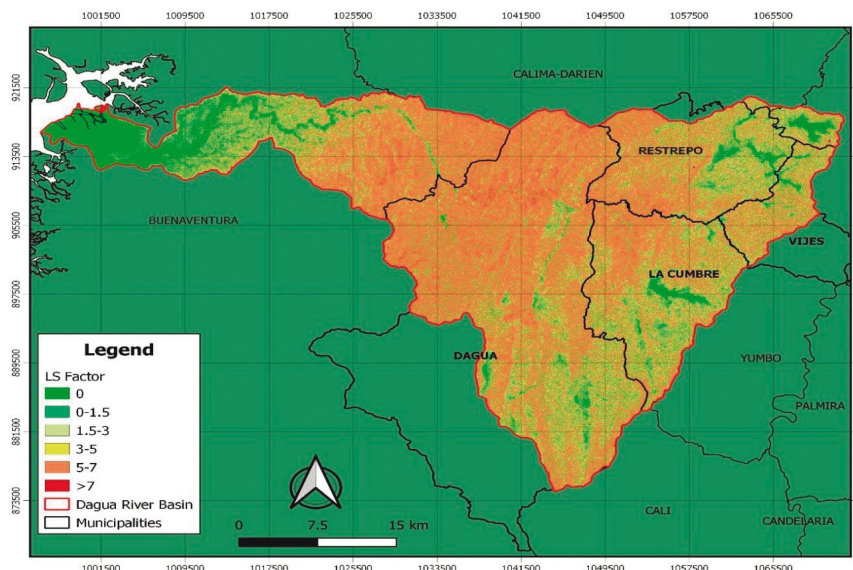


Figure 4. LS factor in the Dagua river basin

3.4 Coverages (Factor C)

It was found that the most representative covers are dense high-ground mixed forest and cultivated pastures, which represent 22.2% and 19.7% of the total area, respectively. However, about crops, the predominant one is pineapple, which covers 1.92% of the area, followed by coffee, banana, sugar cane, and tomato, crops that constitute 1.62% of the total area.

The factor C map is shown in Figure 5; the value ranges from 0.01 to 1.0. The highest C was assigned in the areas without vegetation cover and in the pineapple crop (1.0), taking into account that it is managed as a clean crop and that for at least one year, the fields do not have sufficient plant cover (Sughara, 2001), which can lead to serious erosion problems. The lowest C value corresponds to wooded areas (0.01), pastures (0.04), and crops with covers (0.18-0.5).

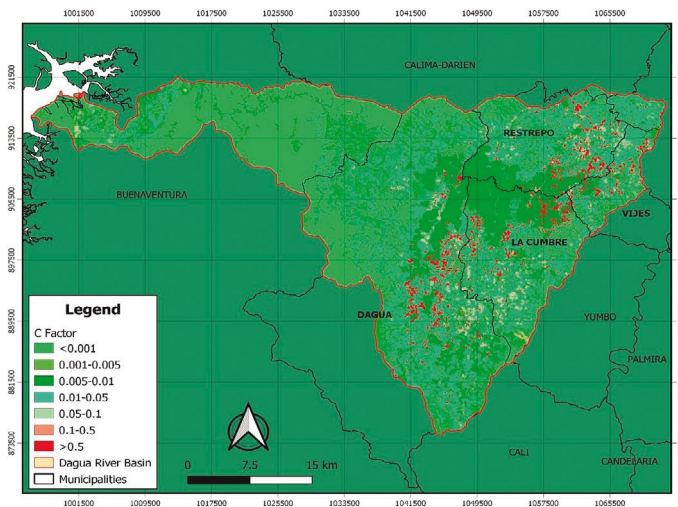


Figure 5. C factor in the Dagua river basin

3.5 Estimation of soil losses in the Dagua river basin

In this regard, it was found that more than a fifth of the area of each of the municipalities of Dagua, Restrepo, La Cumbre, and Vijes, presents an erosion between 100-300 t ha⁻¹ yr⁻¹, where these percentages correspond to 21.22 %, 29.31%, 26.66%, and 25.93%, respectively. These values are framed within the categories of high and very high erosion, according to FAO-UNEP-UNESCO (1980). On the other hand, it was found that the factors that most influenced soil losses were LS and C. This is explained because they are the ones that determine to a greater extent the spatial variations of erosion rates within the same basin or region, where factor c is usually responsible to a large extent for soil losses, and LS is the most determining factor in the transport of the generated sediments (Coleman and Scatena, 1986).

Likewise, Dagua is the municipality with the largest surface area with very high erosion with values that are in the range of 300-500 t ha yr⁻¹ and even higher. The soil loss map is presented in Figure 6.

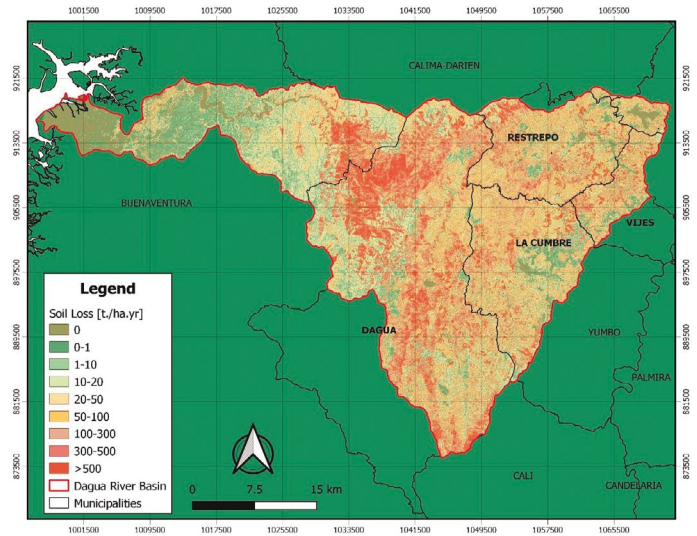


Figure 6. Soil loss Map in the Dagua river basin

On the other hand, pineapple cultivation together with some areas of bare soil are the areas that contribute the most to soil loss due to erosion. In this way, 27.68% of bare soil and 69.24% of pineapple contribute to an erosion of more than 1000 t ha yr⁻¹. Consequently, gullies have been evidenced in the Dagua river basin, which are located mainly on the border between the municipalities of Restrepo and La Cumbre (Rio Grande canyon).

In this regard, Rojas and Campo (2018) mention that erosion processes in this area are predominant, where the highest degrees of erosion are associated with the upper part of the basin, encompassing the municipalities of Dagua, Restrepo, La Cumbre, and Vijes.

4. Conclusions

More than 20% of each of the municipalities of Dagua, Restrepo, La Cumbre, and Vijes presented high and very high erosion categories. Despite the low contribution generated by the R and K factors to the water erosion of the basin, there was a strong influence by the LS factor as the most determining aspect in the transport of the generated sediments, since they are steep slopes, and factor C, which, given the scarce vegetation cover, is largely responsible for these losses.

On the other hand, areas with bare soils and crops such as pineapples contribute to erosion, sometimes exceeding 1000 t ha⁻¹ year⁻¹, an aspect that highlights the importance of plant cover in the prevention and control of water erosion.

It is important to promote soil conservation practices such as planting crops on contour lines, integrated crop cover management, land uses that integrate trees, eliminate inappropriate practices such as widespread burning and weeding, and in more critical cases considering processes of ecological restoration.

References

- Abdulkareem, J., Pradhan, B., Sulaiman, W., & Jamil, N. (2019). Prediction of spatial soil loss impacted by long-term land-use/land-cover change in a tropical watershed. *Geoscience Frontiers*, 10(2), 389-403. <https://doi.org/10.1016/j.gsf.2017.10.010>
- Alarcón, S & Reyes, A. (2013). Erodibility Assessment for Typic Dystrudepts, Typic Hapludands, and Andic dystrudepts by using a rain simulator in the watershed La Centella (Dagua- Valle del Cauca). *Environmental and Natural Resources Engineering*, 12, 49-57.
- Alewell, C., Borrelli, P., Meusburger, K., Panagos, P. (2019). Using the USLE: Chances, challenges, and limitations of soil erosion modeling. *International Soil and Water Conservation Research*, 7 (3), 203-225. DOI: 10.1016/j.iswcr.2019.05.004
- Amundson, R., Berhe, A. A., Hopmans, J. W., Olson, C., Sztein, A. E., & Sparks, D. L. (2015). Soil and human security in the 21st century. *Science*, 348(6235). DOI: 10.1126/science.1261071
- Anache, J. A., Wendland, E. C., Oliveira, P. T., Flanagan, D. C., & Nearing, M. A. (2017). Runoff and soil erosion plot-scale studies under natural rainfall: A meta-analysis of the Brazilian experience. *Catena*, 152, 29-39. <https://doi.org/10.1016/j.catena.2017.01.003>
- Aguirre, M. A., López-Ibarra, L. I., Bolaños-Trochez, F. V., González-Guevara, D. F., & Buitrago-Bermúdez, O. (2017). Percepción del paisaje, agua y ecosistemas en la cuenca del río Dagua, Valle del Cauca, Colombia. *Perspectiva Geográfica*, 22(1), 109-126. <https://doi.org/10.19053/01233769.5402>
- Bennett, H. H. (1939). A permanent loss to New England: Soil erosion resulting from the Hurricane. *Geographical Review*, 29(2), 196-204. <https://doi.org/10.2307/209942>
- Borrelli, P., Robinson, D. A., Fleischer, L. R., Lugato, E., Ballabio, C., Alewell, C., Meusburger, K., Modugno, S., Schutt, B., Ferro, V., Bagarello, V., Van Oost, K., Montanarella, L & Panagos, P. (2017). An assessment of the global impact of 21st-century land use change on soil erosion. *Nature Communications*, 8(1), 2013. <https://doi.org/10.1038/s41467-017-02142-7>
- Borrelli, P., Robinson, D. A., Panagos, P., Lugato, E., Yang, J. E., Alewell, C., Wuepper, D., Montanarella, L & Ballabio, C. (2020). Land use and climate change impact on global soil erosion by water (2015–2070). *PNAS*, 117(36), 21994-22001. <https://doi.org/10.1073/pnas.2001403117>
- Bouma, J., & Montanarella, L. (2016). Facing policy challenges with inter-and transdisciplinary soil research focused on the UN Sustainable Development Goals. *Soil*, 2(2), 135–145. <https://doi.org/10.5194/soil-2-135-2016>
- Cardona, F., Ávila, A. J., Carvajal, Y., & Jiménez, H. (2014). Tendencias en las series de precipitación en dos cuencas torrenciales andinas del Valle del Cauca (Colombia). *Tecnológicas*, 17(32), 85-95.
- Castro, A. F., Lince, L. A. & Riaño, O. (2017). Determination of the risk to the potential erosion by water in the coffee zone of the Quindío, Colombia. *Agricultural and Environmental Research Magazine*, 8(1), 17-26. <https://doi.org/10.22490/21456453.1828>
- Coleman, D. J., & Scatena, F. N. (1986). *Identification and evaluation of sediment sources*. In: Proceedings of a symposium held in Albuquerque, New México, USA, 4-8, August 1986, IAHS Publications, 159, 3-18.
- Corporación Autónoma Regional del Valle del Cauca (CVC). (2007). *Balance Oferta-Demanda de Agua Cuenca del Rio Dagua*. Available online at https://www.cvc.gov.co/sites/default/files/2018-09/Balance_Dagua_0.pdf. (Verified on September 14, 2021).
- Crosson, P. (1995). Soil erosion estimates and costs. *Science*, 269(5223), 461-464. DOI: 10.1126/science.269.5223.461
- Daza, M. C., Reyes, A., Loaiza, W. & Fajardo, M. P. (2012). Índice de sostenibilidad del recurso hídrico agrícola para la definición de estrategias sostenibles y competitivas en la Microcuenca Centella Dagua – Valle del Cauca. *Gestión y Ambiente*, 15(2), 47-58.
- El Kateb, H., Zhang, H., Zhang, P., & Mosandl, R. (2013). Soil erosion and surface runoff on different vegetation covers and slope gradients: A field experiment in Southern Shaanxi Province, China. *Catena*, 105, 1-10. <https://doi.org/10.1016/j.catena.2012.12.012>
- Food and Agriculture Organization of the United Nations (FAO), United Nations Environment Program (UNEP), United Nations Environment Organization (UNESCO) (1980). *Provisional methodology for evaluating soil degradation*. Rome, Italy. 86.
- Gachene, C. K. K., Nyawade, S. O., & Karanja, N. N (2020). Soil and water conservation: An overview. In: Leal Filho, W., Azul, A. M., Brandli, L., Özuyar, P. G., Wall, T. (Eds.) *Zero Hunger. Encyclopedia of the UN Sustainable Development Goals*. Springer, Cham, 1-15.
- González, N., Carvajal, Y. & Loaiza, W. (2016). Analysis of meteorological drought for Dagua river basin, Valle del Cauca, Colombia. *Tecnura*, 20(48), 101-113.
- Guauque, D. E., Rogério de Mello, C., & Curi, N. (2021). Environmental degradation risk by water erosion in a water producer Colombian Andes basin. *Ciencia e Agroecología*, 45. <https://doi.org/10.1590/1413-7054202145010021>
- Instituto de Investigación de Recursos Biológicos Alexander von Humboldt. (2021). *Bosques secos tropicales en Colombia*. <http://www.humboldt.org.co/en/research/projects/developing-projects/item/158-bosques-secos-tropicales-en-colombia>.
- IGAC. (2014). *Instructivo. Códigos para los levantamientos de suelos*. 92p. Available in: <http://igacnet2.igac.gov.co/intranet/UserFiles/File/procedimientos/instructivos/I40100-06-14.V1Codigos%20para%20los%20levantamientos%20de%20suelos.pdf>
- Iroumé, A., Carey, P., Bronstert, A., Huber, A., & Palacios, H. (2011). GIS application of USLE and MUSLE to estimate erosion and suspended sediment loading experimental catchments, Valdivia, Chile. *Revista Técnica de la Facultad de Ingeniería Universidad del Zulia*, 34(2), 119-128.
- Keesstra, S. D., Bouma, J., Wallinga, J., Tittonell, P., Smith, P., Cerdà, A., Montanarella, L., Quinton, J. N., Pachepsky, Y., Van der Putten, W. H., Bardgett, R. D., Moolenaar, S., Mol, G., Jansen, B., & Fresco, L. O. (2016). The significance of soils and soil science towards the realization of the United Nations Sustainable Development Goals. *Soil*, 2, 111-128. <https://doi.org/10.5194/soil-2-111-2016>
- Kirkby, M. J., & Morgan, R. P. C. (1980). *Soil Erosion*. 1st ed, John Wiley and Sons, Chichester, UK, 306 pp.
- Loaiza, W., Reyes, A., & Carvajal, Y. (2012). Application of a Sustainability Index of Water Resources in Agriculture (ISRHA), to define sustainable technological strategies in the Centella watershed. *Ingeniería y Desarrollo*, 30(2), 160-181.
- Loaiza, W., Carvajal, Y., & Ávila, J. A. (2014). Agroecological evaluation of agricultural production systems in the Centella watershed (Dagua, Colombia). *Colombia Forestal*, 17(2), 161-179. <https://doi.org/10.14483/udistrital.jour.colomb.for.2014.2.a03>
- Ma, X., Zhao, C., & Zhu, J. (2021). Aggravated risk of soil erosion with global warming – A global meta-analysis. *Catena*, 200, 105129. <https://doi.org/10.1016/j.catena.2020.105129>
- Meinen, B. U., & Robinson, D. T. (2021). Agricultural erosion modelling: Evaluating USLE and WEPP field-scale erosion estimates using UAV time-series data. *Environmental Modelling & Software*, 137, 104962. <https://doi.org/10.1016/j.envsoft.2021.104962>
- Mitasova, H., Hofierka, J., Zlocha, M., & Iverson, L. R. (1996). Modeling topographic potential for erosion and deposition using GIS. *International Journal of Geographical Information Science*, 10(5), 629-641. <https://doi.org/10.1080/02693799608902101>
- Němetová, Z., Honek, D., Kohnová, S., Hlavčová, K., Šulc, M., Sočuvka, V., & Velísková, Y. (2020). Validation of the EROSION-3D Model through Measured Bathymetric Sediments. *Water*, 12(4), 1-15. <https://doi.org/10.3390/w12041082>

- Núñez, D., Treviño, E. J., Reyes, V. M., Muñoz, C. A., Aguirre, O.A & Jiménez, J. (2014). Using regression models for spatially interpolated monthly average rainfall in the Conchos River Basin. *Revista Mexicana de Ciencias Agrícolas*, 5(2), 201-213.
- Oldeman, L. R., Hakkeling, R., & Sombroek, W. G. (2017). *World map of the status of human-induced soil degradation: an explanatory note*. International Soil Reference and Information Center.
- Pacheco, H. A., Mendez, W., & Moro, A. (2019). Soil erosion risk zoning in the Ecuadorian coastal region using geotechnological tools. *Earth Sciences Research Journal*, 23(4), 293-302. <https://doi.org/10.15446/esrj.v23n4.71706>
- Panagos, P., Standardi, G., Borrelli, P., Lugato, E., Montanarella, L., & Bosello, F. (2018). Cost of agricultural productivity loss due to soil erosion in the European Union: From direct cost evaluation approaches to the use of macroeconomic models. *Land Degradation & Development*, 29(3), 471-484. DOI: 10.1002/ldr.2879
- Pandey, S., Kumar, P., Zlatic, M., Nautiyal, R., & Panwar, V. P. (2021). Recent advances in assessment of soil erosion vulnerability in a watershed. *International Soil and Water Conservation Research*, 9(3), 305-318. <https://doi.org/10.1016/j.iswcr.2021.03.001>
- Perez, E., Suarez, C., Rios, C & Sanchez, J. (1995). *Map of the Potential Natural Environment of the Island of Gran Canarias*. Island Council of Gran Canarias. The Palms. 21-60.
- Ramirez, F., Hincapie, E., & Sadeghian, S. (2009). Erodabilidad de los suelos de la zona central cafetera del departamento de Caldas. *Cenicafé*, 60(1), 58-71.
- Reina-Rodríguez, G. A., Rubiano, J. E., Castro Llanos, F. A., & Otero, J. T. (2016). Spatial distribution of dry forest orchids in the Cauca River Valley and Dagua Canyon: Towards a conservation strategy to climate change. *Journal for Nature Conservation*, 30, 32-43. <https://doi.org/10.1016/j.jnc.2016.01.004>
- Reyes, A., Barroso, F., & Carvajal, Y. (2010). *Guía básica para la caracterización morfológica de cuencas hidrográficas*. Cali: Universidad del Valle. 1-90.
- Rivera, J. H., & Gomez, A. A. (1991). Erosividad de las lluvias en la zona cafetera central colombiana. (Caldas, Quindío y Risaralda). *Cenicafé*. 42 (2), 37-52.
- Riquelme, R., Darrozes, J., Maire, E., Hérial, G., & Soula, J. C. (2008). Long-term denudation rates from the Central Andes (Chile) estimated from a Digital Elevation Model using the Black Top Hat function and Inverse Distance Weighting: implications for the Neogene climate of the Atacama Desert. *Revista Geológica de Chile*, 35(1), 105-121.
- Riquetti, N. B., Mello, C. R., Beskow, S., & Viola, M. R. (2020). Rainfall erosivity in South America: Current patterns and future perspectives. *Science of The Total Environment*, 724, 138315. <https://doi.org/10.1016/j.scitotenv.2020.138315>
- Rojas, A. G., & Campo, L. P. (2018). Assessment of the water quality through the vision of social agents in the watershed of the Dagua River. *Entorno Geográfico*, 16, 50-77.
- Sharda, V., Mandal, D., & Dogra, P. (2021). Prioritizing soil conservation measures based on water erosion risk and production and bio-energy losses in peninsular South Indian states. *CATENA*, 202, 105263. <https://doi.org/10.1016/j.catena.2021.105263>
- US Department of Agriculture (USDA). (1962). *Soil Survey. Soil conservation service in cooperation with California Agricultural Experiment Station*. Seventh. ed, California.
- Wischmeier, W., & Smith, D. (1978). *Predicting rainfall erosion losses: A guide to conservation planning*. United States Department of Agriculture, Agriculture handbook N ° 537, 58.